

Household water treatment and safe storage – effectiveness and economics

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affordable and sustainable manner (Duke et al., 2006) while being simple and easy to manage by their users (Heinsbroek and Peters, 2014). As such, these systems are crucial in reducing occurrence of diarrheal and other debilitating illnesses (Meierhofer and Landolt, 2009; Stauber et al., 2009). Efficiency in providing safe water differs per method. To indicate removal efficiency, the WHO produced guidelines (WHO, 2004) to define default performance targets to indicate a certain removal efficiency for different pathogens as “interim”, “protective” or “highly protective” (see Fig. 1).

$$\text{Price [EURm}^{-3}\text{]} = \frac{\text{Investment} + \text{operational costs}}{\text{Produced water}} \quad (1)$$

When looking at the economics of HWTS systems, it is common practise to look at the price per produced m³ water (NWP, 2010). Generally, this is calculated by dividing the investment and operational costs over the produced water during the lifetime of the technology (NWP, 2010).

The objective of this paper is to give an overview of the potential effectiveness according to the WHO performance targets and the costs paid by the user of three HWTS systems: SODIS, ceramic filters and biosand filters. These HWTS systems are selected because they are widely used, they are promising for the future and sufficient academic literature is written to research their removal efficiency.

1.1 SODIS

SODIS is based on the principle of disinfection by solar radiation (see Fig. 2). The procedure is extremely simple; an unscratched and uncoloured PET or glass bottle is filled with water and exposed to direct sunlight for a minimum of 6 h (Heinsbroek and Peters, 2014). Water with low oxygen and high turbidity levels has to be pre-treated (Acra et al., 1990; Meierhofer and Landolt, 2009).

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1.2 Ceramic filters

The ceramic filter is based on the following principle: a porous media of fired clay that removes microbes by size exclusion and tortuosity (ratio of shortest linear distance over actual flow path) (Hunter, 2009; Sobsey et al., 2008; van der Laan et al., 2014).

- 5 Many variations of ceramic filters exist; e.g. pot filters or “water purifier” (see Fig. 3) (Akvopedia, 2014b; Potters for Peace, 2014), candle filters (Sobsey et al., 2008) and Tulip siphon filter (Basic Water Needs, 2014; Tulipfilter, 2013). Periodic scrubbing and rinsing is necessary to remove impurities (Sobsey et al., 2008).

1.3 Biosand filter

- 10 Biosand filters consist of a concrete or plastic frame filled with crushed rock (sand) filter media of 0.15–0.35 mm particle (Murphy et al., 2010b) (see Fig. 4). Two filter mechanisms govern the removal principle of biosand filters: physical removal of organic matter and turbidity (Sobsey et al., 2008) and biological removal of colloidal particles and harmful pathogens in the so-called Schmutzdeke (Duke et al., 2006; Hunter, 2009; Weber-Shirk and Dick, 1997). The filter can be cleaned manually by removing the top
15 few centimetres of sand and disposing the overlying water (Sobsey et al., 2008).

2 Effectiveness

- In this section an in-depth description is given of the removal mechanisms of each of the selected HWTs systems and the corresponding removal efficiency. Both lab and
20 field studies are used to give an overview of the reported effectiveness. Insufficient data was found on the log removal of protozoa, so this pathogen is excluded from this study.

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and functional proteins), structural changes in cell membrane and reaction with nucleic acids (Russell et al., 1994). There are different ways to impregnate silver in the filter: dipping, painting, pulse injections and fire-in (Oyanedel-Craver and Smith, 2007; Ren and Smith, 2013). Van der Laan (2013) and Oyanedel (2008) did not find a significant difference in removal efficiency for different silver application methods. Neither does an addition of iron appear to increase the removal efficiency in their research (Brown et al., 2008). On the contrary, the storage time in the receptacle of the filter was found to be an important parameter in the bacterial removal efficiency (van der Laan et al., 2014). Concerns exist about the virus removal of ceramic filters, since reported removal efficiencies do not reach WHO guidelines (Murphy et al., 2010a; van der Laan et al., 2014), and show high dispersion (Bielefeldt et al., 2010). No critical parameter was yet identified to improve the virus removal efficiency (van der Laan et al., 2014).

In Figs. 7 and 8, a summary is given of the found removal efficiencies of ceramic filters for bacteria and viruses respectively. The log removal of ceramic filters for bacteria and viruses were between 0.5–7 and 0.09–2.4 respectively based on: Basic Water Needs (2014); Bielefeldt et al. (2010); Bloem et al. (2009); Brown et al. (2008); Brown and Sobsey (2010); Lantagne (2001); Murphy et al. (2010a); Oyanedel-Craver and Smith (2007); Potters for Peace (2014); Roberts (2003); Simonis and Basson (2011); Tulipfilter (2013); van der Laan et al. (2014); Van Halem et al. (2007). For removal of bacteria, most information sources report log removals between 1.3 and 4. The log removals of viruses are significantly lower with most information sources reporting log removals between 0.4 and 1.4.

2.3 Biosand filters

The removal mechanism of the biosand filter is based on the slow sand filtration principle and depends on the daily volume charged to the filter (Elliott et al., 2008). The optimal charge volume is investigated to be equal or smaller than the pore volume (Elliott et al., 2011). When larger charge volumes are exposed to the filter, a decrease in removal efficiency is found (Baumgartner et al., 2007). Although this HWTS system is

ference between safe or unsafe water produced with this filter. Hence, the question arises whether certain removal efficiency can be guaranteed for the HWTS systems.

In Figs. 11 and 12, the total range of lowest to highest log removal reported is shown per HWTS system. It is assumed that every reported removal efficiency has the same likelihood to occur. It can be seen that SODIS have the highest reported efficiency for bacteria removal (9 log removal) whereas, biosand filter report the highest reported efficiency for virus removal (7 log removal). Biosand filters show the lowest (zero) removal efficiency for bacteria whereas for virus removal, all filters have been reported with a zero log removal in one or more studies.

2.5 Critical evaluation of effectiveness

Recent studies put critical notes to the effectiveness of HWTS systems. Results of field tests reveal that HWTS systems do not always improve and sometimes even worsen the pathogenic state of the influent (Murphy et al., 2010a). The lack of blinding and considerable heterogeneity in the results of HWTS systems show signs of concerns (Hunter, 2009). Moreover, it is reported that the research method can have a big impact on the reported efficiency (van der Laan et al., 2014). The reported removal efficiency also depends on the indicator pathogen used, as shown by Palmateer (1999) and Elliot (2008). Quality tests are not yet globally standardized (Rayner et al., 2013), so that a fair comparison between data sets is challenging.

2.6 Human factors

Research shows that operating conditions can reduce the effectiveness of HWTS systems (Baumgartner et al., 2007). The effectiveness of HWTS systems does not only depend on technology, but also on human factors. When the HWTS system is not operated properly, exposure to pathogens can remain high. For example, it is common that people use the storage container of the device to collect dirty untreated water to feed the HWTS system, reducing the effectiveness of the device (Murphy et al.,

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2010a). Other reasons why in practice the effectiveness of HWTS systems is reduced: (i) people only treat part of their used water (Sobsey et al., 2008), as the water supply of HWTS systems can be reduced in time due to clogging (ii) people are unable to purchase replacements (Brown et al., 2009; Hunter, 2009; Meierhofer and Landolt, 2009), (iii) people only treat their water intermittently (Sobsey et al., 2008), (iv) people have limited guidance to determine whether pre-treatment is necessary (Sobsey et al., 2008), (v) people have simply stopped using the device (Hunter, 2009), (vi) or have sold it to a friend or relative (Brown et al., 2009). For ceramic filters, the rate of participation reduction is estimated at $2\% \text{ month}^{-1}$ (Brown et al., 2009). The found diversity in effectiveness prompts that sufficient training and continued monitoring is needed to increase and sustain proper HWTS device management. Preferably, this could be done by a well-embedded local agent in order to increase acceptability (Meierhofer, 2006). Understanding the human factors that influence the real effectiveness of the HWTS systems is crucial for widespread adoption and sustained usage (Sobsey et al., 2008).

3 Economical evaluation

In this section, the parameters that determine the purchase price of a HWTS system (see Fig. 12) and the reported prices for the three selected HWTS systems are discussed (see Table 1).

3.1 Economic Parameters

The price of HWTS systems depends strongly on project area (K. Wagoner, personal communication, 2014; H. Jansen, personal communication, 2014). This can be explained by the fact that the price of HWTS systems is determined by (at least) 5 parameters (see Fig. 12). The first parameter of the costs for HWTS systems is the production costs (or investment costs) (Basic Water Needs, personal communication, 2014; K. Wagoner, personal communication, 2014). These costs depend on the type

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of HWTS system and the region of production. Factories in China and India are frequently used, due to lower labour costs. The second parameter is distribution (Stuurman et al., 2010). Transport costs between production and project area depend on quantity and weight. In-land and over-sea transportation can differ significantly in total cost (Basic Water Needs, personal communication, 2014). This parameter is estimated to be the most dominant (Basic Water Needs, personal communication, 2014). However, transportation costs are often not included in the reported price values on websites or in literature, because of the high variability per project area. Local production factories are established to diminish distribution costs and enhance local economy (Brown, 2007). The third parameter is taxes. Depending on the country, HWTS systems need to be imported and import fees are involved. These costs can be relatively high (Basic Water Needs, personal communication, 2014). A possible fourth parameter is whether subsidies are present for the project (Stuurman et al., 2010). Often, NGOs arrange donor projects where consumers pay a reduced price (Stuurman et al., 2010). However, research points out that consumers use the provided filters more, when they really have to invest to purchase the technology (Brown et al., 2009). A possible fifth parameter is the (local) distributor's marginal fee needed to maintain the business (Basic Water Needs, personal communication, 2014). Depending how the supply chain is constructed, a (local) distributor organizes the sales in the project area. Furthermore, it should be mentioned that the price is time-dependent and susceptible to exchange rates.

3.2 Costs of HWTS systems

A limited number of peer-reviewed articles mention costs of HWTS systems, and only retail prices were mentioned. Retail price depend on the 5 parameters mentioned in the previous section and is the price eventually paid by the user. The retail price could be converted to the price per m^3 , when the potential amount of water that can be treated with one filter is known (see Fig. 13). This potential amount is dependent on the lifetime of the filter, the flowrate, the sustainability of the system, e.g. Reliable and

sufficient information about this total potential amount is lacking. Therefore the retail price is mentioned separately from the price per m^3 water produced (see Table 1).

In Table 1, a summary is given of the price per m^3 and retail price per HWTS system. SODIS has a retail price of USD 0, since old PET bottles can be used. When new bottles are used, only a small investment is necessary (NWP, 2010). The costs per m^3 are related with the retail price. The outlier of $3\text{--}8\text{ m}^{-3}$ by (Akvopedia, 2013) is assumed to be an error, since it does not correspond with the numbers in the rest of the text of the same reference. For ceramic filters, the range of retail prices is between USD 1.9–30, with most of the references mentioning prices around USD 15. The differences in price can be explained by the parameters elaborated in the previous section. The price per m^3 ranges between USD 0.3 and USD 5.2, which depends on the estimation of the potential amount of water that can be treated with the filter. For biosand filters, large ranges are found in the price per m^3 and retail price: EUR 0.07–10 and USD 7–100 respectively. The outlier in price per m^3 of USD 10 is unreliable, since no argumentation is given in the reference (Akvopedia, 2014c). The outlier in retail price of USD 100 for concrete is also stated without further explanation (Sobsey et al., 2008).

Overall, it is found that the biosand filter has the lowest price per m^3 produced what can be explained by its long life time, low maintenance costs and sustainable flowrate. Biosand filters do have the highest retail price (even after the highest outlier is neglected). Although SODIS is also a cheap technology, it requires a (small) investment when new (glass) bottles are used and it only produces little amount of water per bottle. SODIS does have the cheapest retail price. It is shown that ceramic filters have the biggest range of price per m^3 water with the highest numbers. Ceramic filters are prone to breakage and the flowrate can decrease over time due to clogging. By far most independent research exists on ceramic filters compared to the other HWTS systems.

In Figs. 14 and 15, an overview of the price ranges is given, neglecting the outliers mentioned above.

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3.3 Constraints to economic evaluation

Most cost estimations of HWTS systems are found on websites of coordinating NGOs or device suppliers. Because the information is practice-oriented, the reliability of this information is likely to be fluctuating. More direct information from local producers turned out to be necessary. Resource Development International in Cambodia reveals a standing quotation of a ceramic filter for USD 12 (RDI, 2014), which is in line with the prices in other sources. Since the price of HWTS systems does not only depend on the 5 parameters mentioned before, but is also fluctuating in time and susceptible to exchange rates. The price of the HWTS system today is different from the price indicated for 2007 (H. Jansen, personal communication, 2014). This study does not include these changes. The prices mentioned in Table 1 are considered to be valid for the year of the respective reference.

4 Conclusions

In this study the economics and removal efficiencies of three selected Household Water Treatment and safe Storage (HWTS) systems were compared: SODIS, ceramic filters and biosand filters. The costs of HWTS were based on five parameters: production, distribution, taxes, subsidies and marginal fees. They influence the price paid by the consumer besides other factors (interest, inflation). Additionally, the produced volume of water, or lifetime of the HWTS determines the actual price per m³. The reported log removal should be viewed with some precaution, as parameters like indicator pathogen, research method and human factors are of influence. Also, most studies are short-time (around 26 weeks) and designed poorly unblended, which could have given biased results (Hunter, 2009; Sobsey et al., 2008).

For SODIS, low retail prices and intermediate prices per m³ were observed with a range of removal efficiencies for bacteria from “non-protective” to “highly protective” and for viruses from “non-protective” to “protective” according to WHO targets. Ceramic

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filters showed intermediate retail prices and high prices per m³ with a range of removal efficiencies for bacteria from “non-protective” to “highly protective” and for viruses “non-protective”. Biosand filters had high retail prices and low prices per m³ with a range of removal efficiencies for bacteria from “non-protective” to “protective” and for viruses from “non-protective” to “highly protective”. Overall, a relationship between HWTS removal efficiency and economics was not observed.

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Table 1. Overview of the price and retail price of HWTS systems.

Technology	Adjustment	Price (USD m ⁻³)*	Retail Price (USD)*	Reference	
SODIS	–	0.53–2.03 new	0	NWP (2010) NWP (2010)	
		0–0.27 used 3–8	0	akvopedia (2013) CAWST (2012)	
Ceramic	General	0.3–0.61	5–15	NWP (2010)	
			8–10	Oyanedel-Craver and Smith (2007) Brown et al. (2009), Sobsey et al. (2008)	
	Tulip Siphon Filter	0.4–0.44	5.69–7.32	NWP (2010)	
		1.14–5.2 2–5	26.02	Tulipfilter (2013) Basic Water Needs (2014)	
	Pot Filter	1		9.76–17.89	akvopedia (2014b) CAWST (2012)
				7.5	Roberts (2003)
5.4–28				NWP (2010)	
Water4life	0.46		akvopedia (2014a) NWP (2010)		
Candle Filter Potters for Peace			1.87–20.33	NWP (2010)	
			15–30	CAWST (2012)	
Biosand	Concrete	0.07–0.15 10	15–25	CAWST (2012), Potters for Peace (2014b)	
			12	(RDI, 2014)	
			6.99–22.76	NWP (2010)	
	Plastic Iron oxide filter			12–40	akvopedia (2014c), Sswm.info (2014)
				25–100	Sobsey et al. (2008), Ahammed and Davra (2011)
				7–28	CAWST (2012)
			12–30	CAWST (2012)	
			75	Ahammed and Davra (2011)	
			15–36		

* Conversion used where necessary 1.23 EUR/USD (Bloomberg, 2014).

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Target	Log₁₀ reduction required: Bacteria	Log₁₀ reduction required: Viruses	Log₁₀ reduction required: Protozoa
<i>Highly protective</i>	≥ 4	≥ 5	≥ 4
<i>Protective</i>	≥ 2	≥ 3	≥ 2
<i>Interim*</i>	Achieves "protective" target for two classes of pathogens and results in health gains		

Figure 1. WHO guidelines on default performance targets of HWTS systems (WHO, 2004).

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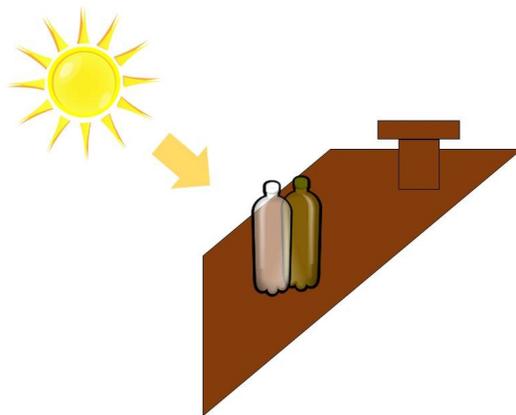


Figure 2. Schematic illustration of SODIS.

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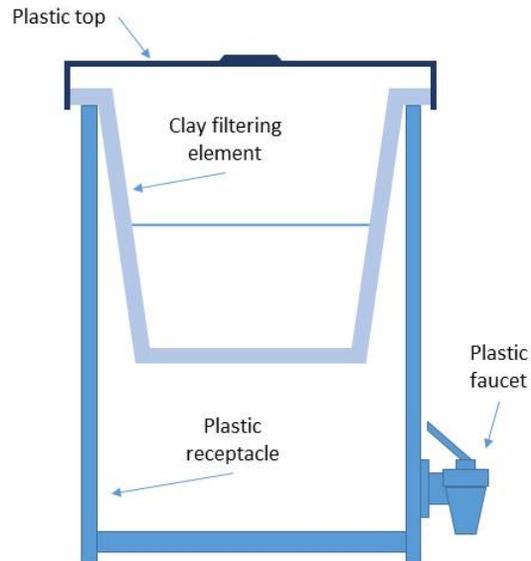


Figure 3. Schematic illustration of ceramic Pot filter (Van Halem et al., 2007).

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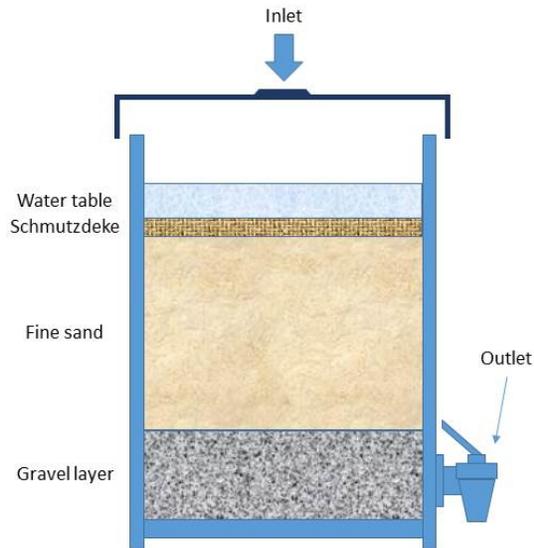


Figure 4. Schematic illustration of biosand filter.

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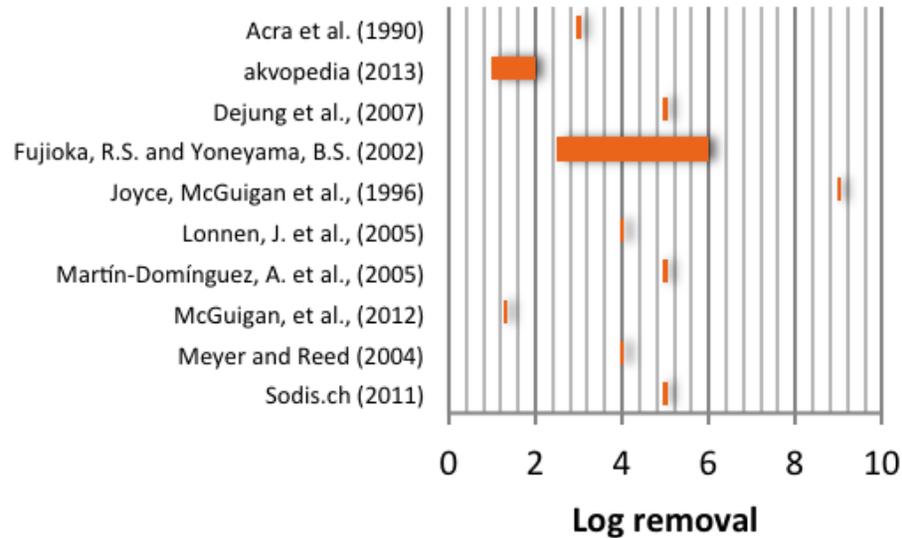


Figure 5. Log removal of SODIS for bacteria.

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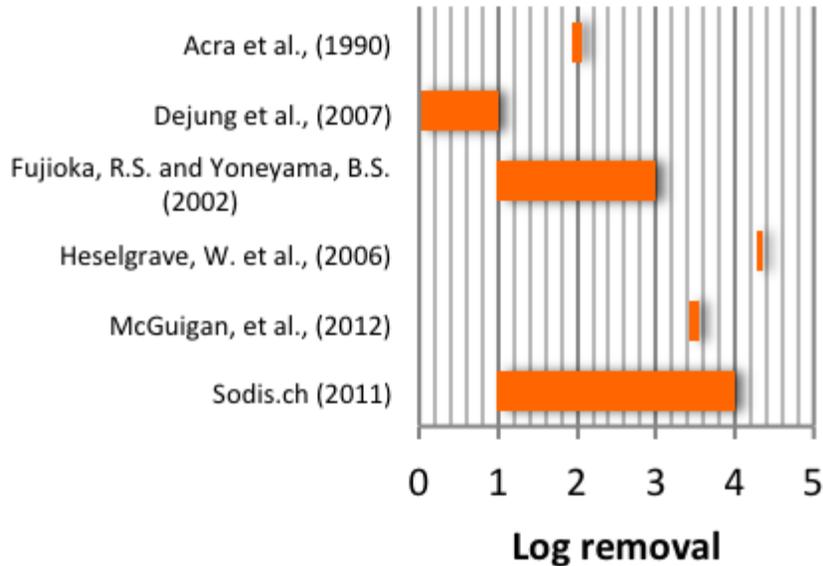
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**Figure 6.** Log removal of SODIS for viruses.[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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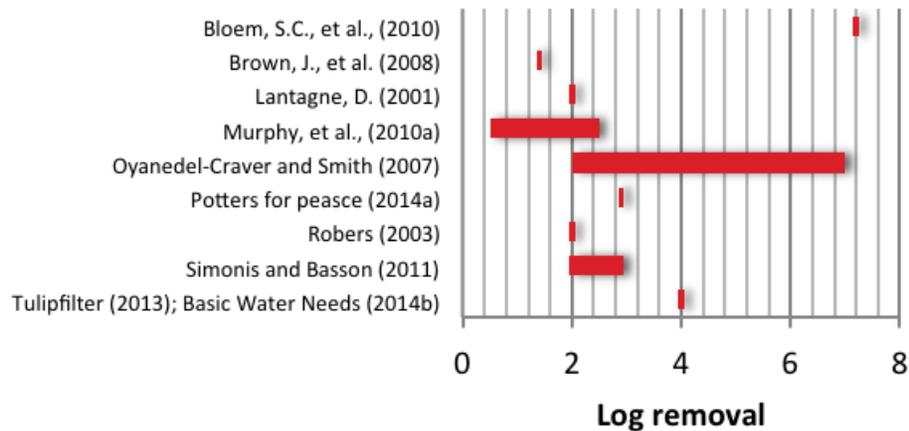


Figure 7. Log removal of ceramic filters for bacteria.

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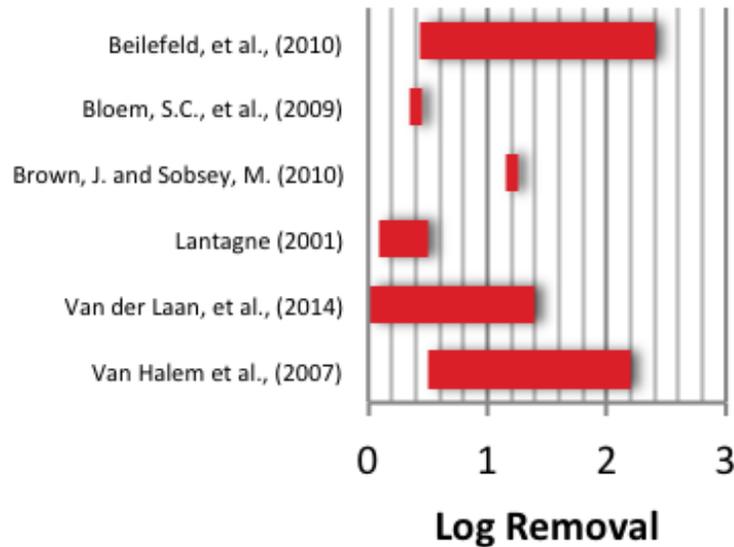
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**Figure 8.** Log removal of ceramic filters for viruses.[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[⏪](#)[⏩](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Printer-friendly Version](#)[Interactive Discussion](#)

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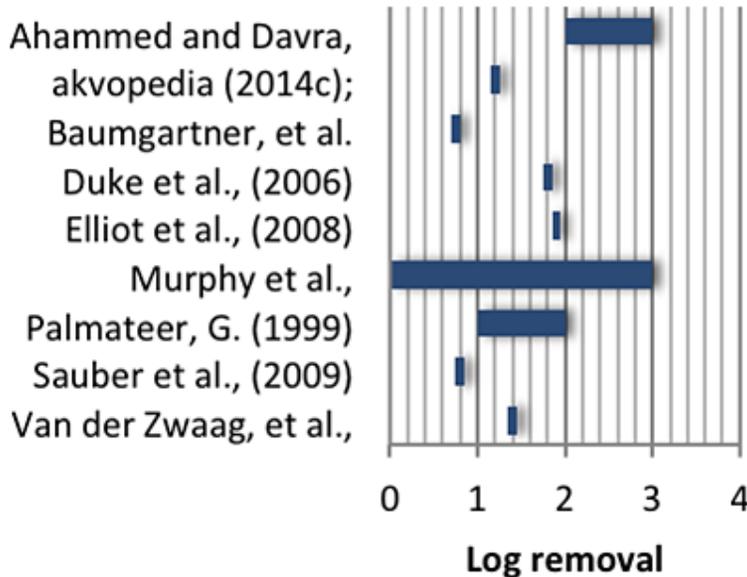


Figure 9. Log removal of biosand filters for bacteria.

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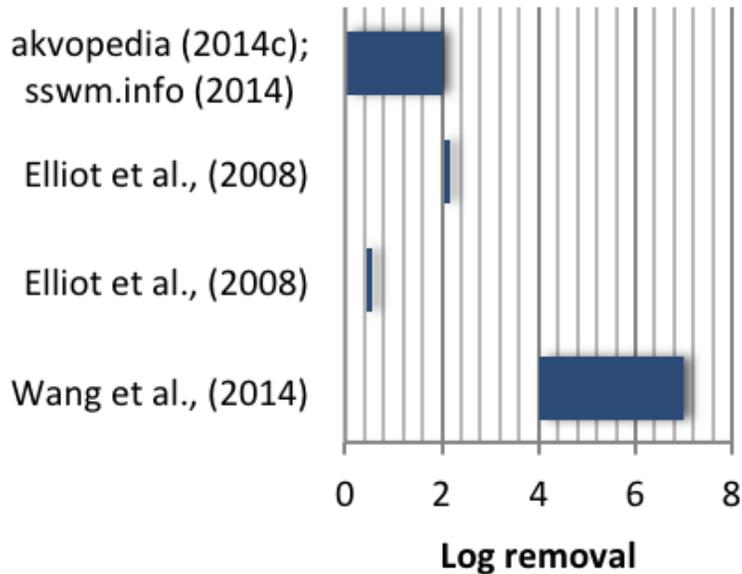


Figure 10. Log removal of biosand filters for viruses.

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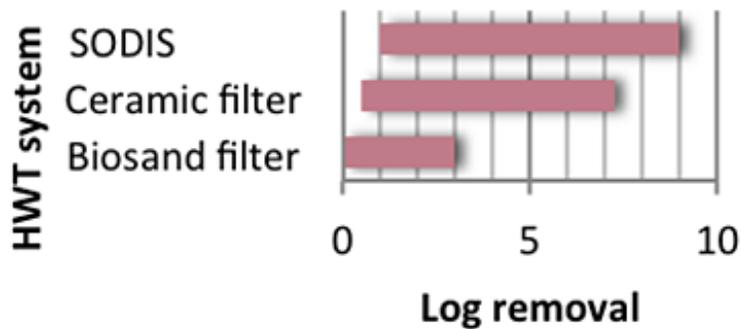


Figure 11. Overview of the overall range of found log removals in Sects. 2.1–2.3 for bacteria.

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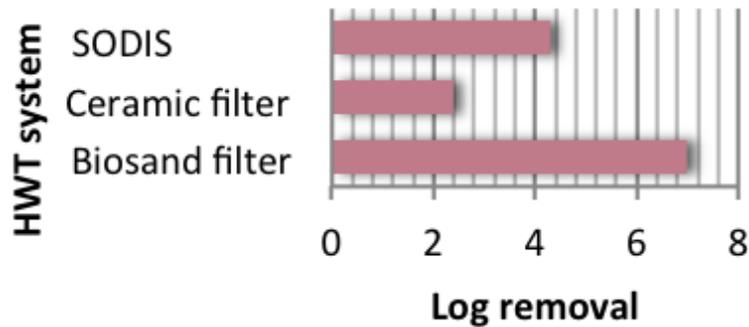


Figure 12. Overview of the overall range of found log removals in Sects. 2.1–2.3 for viruses.

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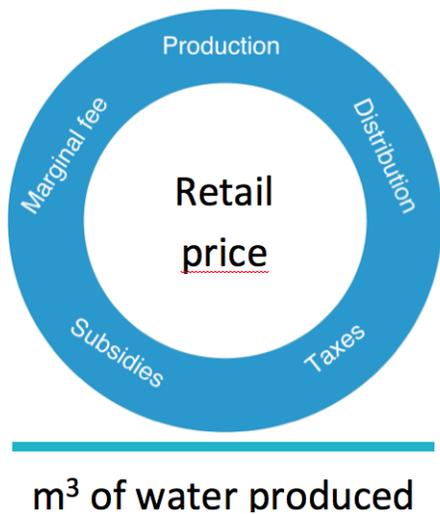
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= price per m³

Figure 13. Parameters that determines the retail price of HWTS systems.

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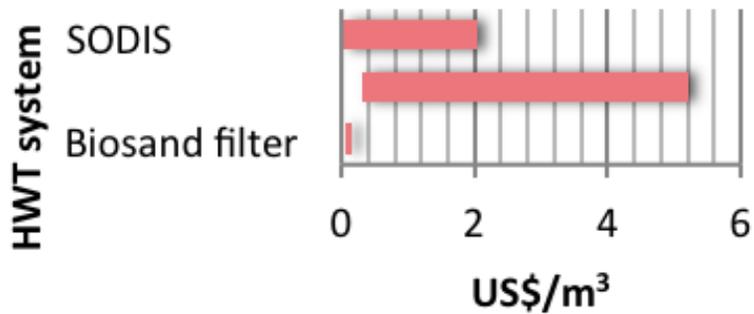


Figure 14. Price ranges (USD m⁻³) for the three selected HWT systems (outliner of biosand filter is neglected).

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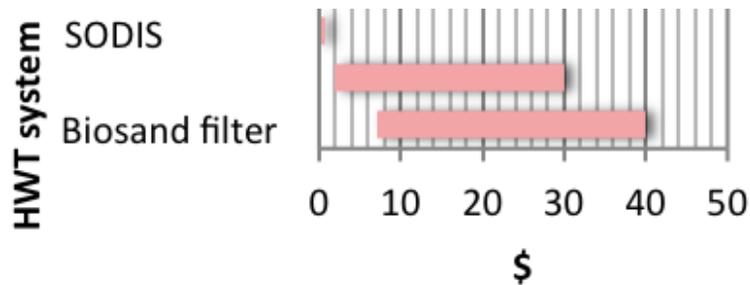


Figure 15. Ranges for the retail price of the three selected HWT systems (outliner of SODIS is neglected).

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